

1 **Particle Number and Size Distribution Emissions During Light-**
2 **Duty Vehicle Cold Start Using the Total On-Board Tailpipe**
3 **Emissions Measurement System**

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1 **ABSTRACT**

2

3 Vehicle emissions during cold start are known to be significantly higher than after optimal
4 vehicle operating temperatures are reached. There is limited data however, on particle number
5 and size distributions during cold start. Cold start tailpipe emissions from a 1999 Toyota Sienna
6 minivan were quantified at ambient temperatures between 20°C and 37°C using a novel system
7 (Total On-board Tailpipe Emissions Measurement System (TOTEMS); [see [1]]) assembled to
8 quantify the full suite of exhaust emissions from light-duty vehicles. TOTEMS particle number
9 distributions were measured from 5.6–562 nanometers using an Engine Exhaust Particle Sizer
10 (EEPS) and total 3–3000 nanometer particle counts were measured using an Ultrafine
11 Condensation Particle Counter (UCPC) with 1-second temporal resolution during cold start and
12 on-road driving.

13

14 Second-by-second particle number distributions from 5 cold start emissions tests showed similar
15 particle emissions patterns, allowing for 3 different cold start Phases to be identified based on
16 particle number emissions behavior. Cold start duration ranged from 165 to 230 seconds and
17 increased with decreasing ambient temperature. Different particle sizes during each Phase were
18 emitted for different lengths of time, with the most abundant particles in the nanoparticle
19 (diameter < 50 nm) range. The mean particle number distributions showed more than 99% of
20 total particle number below 100 nanometers. Concentrations of ultrafine particles (<100
21 nanometers) during cold start were at least 10–100 times (EEPS), and as much as 1000 times
22 (UCPC) above hot-stabilized idle emissions. Observations also suggest the presence of tiny
23 particles below 6 nanometers during cold start.

24

1 INTRODUCTION

2
3 Recent evidence shows a strong relationship between fine particulate matter exposure and
4 increased rates of lung diseases, cardiopulmonary diseases and mortality [2-5], as well as
5 decreased lung development in children [6]. Vehicles are a major source of fine (particle
6 diameter $D_p < 2.5$ micrometers), ultrafine ($D_p < 100\text{nm}$) and nanoparticles ($D_p < 50\text{nm}$).
7 However, vehicle emissions tests, typically performed under idealized lab conditions, are not
8 wholly representative of the variability and emissions levels associated with real-world vehicle
9 operating conditions. This study is part of a larger project initiated to better understand emissions
10 from vehicles under real-world, on-road driving conditions. The larger project (see [1] for more
11 detailed description) is unique because total gas (28 species) and particulate (number
12 distribution) tailpipe emissions are being collected for light-duty gasoline vehicles operating on a
13 specified driving route in Vermont over four seasons of sampling. Here, 1-second data on
14 particle size distributions from 5.6 to 562 nanometers and total particle number from 3-3000
15 nanometers collected from a single vehicle are examined to quantify particle number emissions
16 behavior during cold start and warm-up route driving after engine off times greater than 12
17 hours. A 12-hour engine off time is sufficient to cool the engine and catalyst to ambient
18 temperature [7, 8].

19 Cold start particle number emissions are significantly higher than emissions after a light-
20 duty vehicle has warmed up to its normal operating temperature [9]. However, little real-world
21 data exists on particle number distributions during cold start. Extensive studies have been
22 conducted to quantify cold start gas-phase emissions [8, 10], but these studies are typically,
23 though not always, limited to the criteria pollutants and lack high temporal resolution data for the
24 full particle size distribution, often times focusing on only gases or particle mass. Furthermore,
25 data on particle number most often lacks high temporal resolution for the entire distribution.
26 When operating efficiently, the three-way catalytic converter has been found to remove 75% of
27 hydrocarbon particle precursors under low speed and load operating conditions [9]. For the EPA
28 FTP75 drive cycle, cold start produced more than 80% of the total hydrocarbon gaseous
29 emissions [11], which suggests possible increases may be expected for cold start particle number
30 emissions due to gas-to-particle nucleation upon exhaust cooling. Generally, peak particle
31 emissions also tracked peak emissions of CO, NO_x and hydrocarbons [9]. Understanding the
32 magnitude of these cold start emissions is important because many trips – especially in urban
33 areas – are short and never allow the vehicle to reach normal operating temperatures [10].
34 Furthermore, most trips begin in urban environments, resulting in significantly higher real-world
35 emissions than predicted by existing models.

36 Previous studies have attempted to quantify the cold start size distribution and
37 composition of particle number emissions from light-duty vehicles. Using a direct injection
38 gasoline engine and Cambustion's differential mobility analyzer, particle number emissions were
39 found to decrease over time during engine warm-up, often with particle size decreasing as well
40 [12]. One study noted particles from gasoline engines are primarily in the 7 to 55 nm range
41 during cold start [13]. It was found that only 29 percent of particles were in the solid phase:
42 elemental carbon between 5 and 10 nm and carbonaceous particles in the 30 – 80 nm range [12],
43 meaning substantial nucleation occurred. Now widely accepted, the low concentration of larger
44 solid particles favors the formation of nuclei mode (diameter < 30 nm) particles [14] due to the
45 small surface area available for gas-to-particle condensation. Also, many studies have focused
46 more on diesel engines because they traditionally have had higher emissions rates, but diesel

1 vehicles have lower cold start emissions than gasoline vehicles [15, 16]. It is important to note
2 that in all previous studies, the particle size range of the instruments used did not capture the
3 whole particulate emissions size range, meaning a significant number of particles may not have
4 been quantified.

5 Decreases in ambient air temperature have been linked to increases in exhaust emissions
6 that are an order of magnitude or more above hot-stabilized emissions [15, 17-19]. Ambient
7 temperatures between -20°C and +22°C had a significant effect on criteria pollutants, including
8 NOx, CO, HC, and particle mass [19], although the reported effects on particle mass are
9 inconsistent [15, 19]. One group documented hydrocarbon emissions were 26 grams per start
10 higher at -20°C than at +23°C [16] Another group reported hydrocarbon emissions were reduced
11 by a factor of 4 during cold starts when ambient temperature rose from -2°C to 32°C [18]. These
12 changes in hydrocarbon emissions could have a similar effect on nuclei mode particles due to the
13 relationship between hydrocarbon gas emissions and particle nucleation. Another group found
14 similar results, with cold start emissions many times (and for some pollutants, order of
15 magnitude) higher than hot operation, with this effect magnified as ambient temperature
16 decreased [19]. Reported particle mass increased by a factor of 19 when cold start temperature
17 was decreased from 22°C to -7°C [19]. A more recent study showed that particle mass was less
18 influenced by ambient temperature than the gaseous criteria pollutants [15]. However, this result
19 may have no relation to number emissions as particle number and mass seem to have little
20 correlation [20]. Mass is often times reported because current regulations are mass-based (PM10
21 and PM2.5), but particle size and number have the greatest effect on human health. Therefore,
22 this study focused on particle number from 3 to 3000 nanometers and quantified the distribution
23 from 5.6 to 562 nanometers. Most particles measured in this study – nearly 100% – were below
24 300 nanometers in aerodynamic diameter.

27 **METHODOLOGY**

29 This section describes the vehicle and its operation during cold start tests, the Total On-Board
30 Tailpipe Emissions Measurement System (TOTEMS) instrumentation and protocols relevant to
31 this study and laboratory instrument validation test procedures. All data for this study was
32 collected with a minimum temporal resolution of 1-second.

34 **Test Vehicle and Emissions Data Collection**

36 Individual emissions tests consisted of a single driver operating the vehicle under real-world
37 driving conditions over a specified sequence of startup operations and a warm-up driving route in
38 Burlington, Vermont. One driver was used throughout the testing to minimize inconsistencies
39 resulting from driver behavior. Prior to starting the vehicle, a series of quality assurance/quality
40 control (QA/QC) measurements and operations were performed in order to collect accurate
41 instrument and vehicle baseline data for each run. These QA/QC procedures included 3-hour
42 particle instrument warm-up, collection of instrument blank data, collection of exhaust pipe
43 tunnel blank data and synchronization of all instrument computer clocks via GPS software.

44 A gasoline fueled 1999 Toyota Sienna minivan with a V-6, 194 horsepower, 3.0 liter
45 engine and a 4-speed automatic transmission with LEV 1 emissions certification was used for all
46 tests. At the time of testing, the minivan odometer ranged between 148,424 and 148,702 miles.

1 Although an older vehicle was tested here, a 2005 study reported an 8.9 year median age of
2 passenger cars. This was up from a median of 6.5 years in 1990. Furthermore, between 1990 and
3 2005, the median age of vehicles trended steadily up year-to-year, representing a sustained
4 interest in emissions from these older vehicles [21]. After completion of the blanks and
5 determination that all instrumentation was functioning properly, the test vehicle's engine was
6 started. The minivan then idled in Park for a period of time prior to shifting into Drive and
7 traveling a 0.8 mile 'warm-up' route to a gas station. Three traffic signals and two stop signs
8 were encountered along the this route. For this study, data are presented from engine start to
9 engine stop at the gas station.

10 This activity was more than adequate for complete warm-up during the spring season
11 data collection reported here. During future winter months testing, an additional 2.5 mile urban
12 section with steep hills will be utilized to ensure complete engine warm-up prior to driving the
13 full route.

14 15 **Total On-Board Tailpipe Emissions Measurement System (TOTEMS)**

16
17 As described in more detail elsewhere [1], a suite of 20 instruments and accompanying data
18 acquisition were assembled to collect second-by-second vehicle operation and tailpipe emissions
19 data as test vehicles operate in the real-world. This package, TOTEMS, collects emissions data
20 on particle number, particle number distribution as well as 28 gas species. Here, only the
21 instrumentation and methodology relevant to particle emissions and characterization of cold start
22 vehicle behavior are described. An on-board battery system powered all instruments without
23 drawing electrical power from the test vehicle. An automatic transfer switch was used to transfer
24 on-board instrumentation from grid to battery power without shutting the instruments down,
25 saving battery power while allowing adequate instrument warm-up.

26 Relative humidity and ambient temperature were monitored both in and outside the
27 vehicle using identical HOBOWare pro v2 sensors. An AutoEnginuity ST01 ScanTool collected
28 4 engine parameters; engine load, vehicle speed, mass air flow and engine RPM. Data
29 acquisition cards were used to collect data for the following instruments via a Labview 7.0
30 interface: a type J thermocouple to monitor exhaust temperature at the end of the tailpipe, type T
31 thermocouples to monitor temperature at the end of the heated line and the inlet of the emissions
32 instruments, four differential pressure transducers to quantify exhaust flowrate and four
33 parameters monitoring the exhaust dilution system for particle measurements. The dilution
34 system for particle measurement included two separate components – the Matter Engineering
35 MD19-2E Rotating Disk Mini-diluter and the TSI Air Supply Evaporation Tube (ASET 15-1) –
36 designed to work together providing first stage (MD19-2E) and second stage (ASET 15-1)
37 dilution in one self-contained device. A total dilution ratio of 1:120 (raw exhaust : HEPA filtered
38 dilution air) was used throughout emissions testing.

39 Exhaust was transferred via a heated line from the tailpipe to the particle dilution system.
40 Particle number emissions were quantified with two TSI, Inc. instruments: a Model 3090 Engine
41 Exhaust Particle Sizer Spectrometer (EEPS) and a Model 3025A Ultrafine Condensation Particle
42 Counter (UCPC). The EEPS counts ($\pm 20\%$ accuracy) and sizes ($\pm 10\%$ accuracy) particles
43 across 32 different size ranges from 5.6 to 562 nanometers. The midpoint of each EEPS size
44 range is the reported particle size based on aerodynamic diameter. The EEPS records particle
45 number distribution data at a rate of 10 Hertz, but values were aggregated to a 1 Hertz rate. The 1
46 Hertz measurements are the discrete average of all measurements within a given second and

1 were recorded to the on-board emissions PC using TSI EEPS version 3.1.0 software. The UCPC
2 counts particles from 3–3000 nanometers with a reported detection efficiency of 50% for 3
3 nanometer particles and 90% for particles greater than 5 nanometers. The total particle number
4 concentration from the UCPC was recorded to the on-board emissions PC using Aerosol
5 Instrument Manager Software (version 8.1.0).

6 Anti-vibration platforms were constructed for both the EEPS and UCPC to minimize
7 inaccuracy and instrument error due to noise resulting from vibration while driving. The
8 instrument platform for the UCPC uses 6 natural rubber mounts, and because the UCPC is
9 influenced little by vibrations, these mounts serve to help minimize instrument malfunctions that
10 could result from being jostled. The EEPS is mounted on 10 silicone gel type mounts, which
11 reduced electrometer noise caused by driving by 64%.

12 The TOTEMS emissions measurement setup pulls engine exhaust from the tailpipe
13 adapter through the 10-foot, 3/8-inch ID heated line set to 191 C at an exhaust sample flowrate of
14 13 liters/min (Lpm). The heated line (Atmoseal model IGH-120-S-6/X-G13) is lined with
15 smooth, flexible stainless steel pipe to minimize particle losses. At the end of the heated line is a
16 4-way, stainless steel Swagelok fitting that splits the flow of undiluted exhaust to the emissions
17 instruments: 12 Lpm to the FTIR for gas analysis and 1.0 Lpm to the dilution system for particle
18 measurements. From the dilution system, particles are transported to the EEPS and UCPC via 42
19 inches of 1/4-inch ID conductive silicone tubing. The flow is split, first to the EEPS, then to the
20 UCPC using 2 stainless steel Swagelok fittings. Data is collected on a Dell OptiPlex GX620
21 desktop “Emissions PC” outfitted with two National Instruments model 6024E data acquisition
22 cards and 5 serial ports.

23 EEPS and UCPC particle number emissions data were plotted for analysis using
24 MATLAB version R2008b . Instrument blanks were used in this program to calculate the lower
25 detectable particle concentration limit for each EEPS size range. Because the EEPS operates on
26 electrical principles, it is sensitive to noise. Therefore, emissions values less than three times the
27 standard deviation of the instrument blank data were considered below detection and omitted
28 from the analyses. Also, using the pre-run instrument blank for each run, the mean noise values
29 for each of the 32 EEPS size ranges were calculated individually and subtracted from the EEPS
30 concentrations. Noise quantification for the UCPC was unnecessary as it was not sensitive to
31 noise .

32 **Data Alignment for Instrument and Exhaust Flow Lags**

34 All particulate number emissions data were aligned to the scantool data based on the “ignition
35 on” point for reference. At this point, engine RPM changed from 0 (engine off) to the peak
36 reading (ignition) in one recorded second of data. The EEPS and UCPC behaved similarly,
37 where the concentrations jumped from background to maximum concentration in one recorded
38 second. Therefore, the scantool and particle emissions data were time-aligned using the last point
39 of 0 RPM and the last point for background particulates. No lag was assumed for exhaust
40 temperature because it appeared to be sufficiently small (<0.5 seconds) to not affect data
41 alignment.

1 Particle Number Instrument Comparison: Laboratory Tests

2
3 A comprehensive comparison between the EEPS and UCPC was conducted. From this
4 comparison, it was determined that both particle instruments are necessary for both quality
5 assurance and to fully understand the second-by-second tailpipe particle emissions data.

6 Laboratory tests were performed by generating constant concentrations ($\sim 3.5 \times 10^4 \text{ \#/cm}^3$)
7 and varied concentrations (5.4×10^3 to $8.9 \times 10^4 \text{ \#/cm}^3$) of 10–100 nanometer particles, ensuring
8 no detection bias towards either particle instrument (The EEPS and UCPC have overlapping, but
9 not identical, particle size measurement ranges). Two types of particles were generated: a sodium
10 chloride solution with a concentration of 0.2 g/L in deionized water and a 10:1000 (v/v) emery
11 oil : isopropyl alcohol solution. Both solutions were atomized using a TSI model 9302 atomizer
12 and particle-free compressed air. From the atomizer, the NaCl aerosol passed through silica gel
13 desiccant for drying. The desiccant was bypassed when using emery oil (drying not needed).
14 Aerosol was transferred to the EEPS and UCPC through the dilution system (set to 80°C) using
15 the same setup described for emissions testing. A total of 28 tests using sodium chloride particles
16 and 14 tests using oil particles were conducted. Particle type (emery oil versus sodium chloride)
17 did not influence the total particle concentrations from either instrument.

18 No bias in aerosol flow to either instrument was seen for the setup described above.
19 There was a detectable bias in total concentrations, however, because the EEPS has a much
20 higher maximum concentration limit, which is dependent on the individual size ranges [10^7
21 \#/cm^3 in size range 1 to 10^5 \#/cm^3 in size range 32] and the UCPC frequently reached its
22 maximum concentration [10^5 \#/cm^3] during the cold start. The in-lab comparison confirmed that
23 both instruments have response times to enable analysis of cold start data and that the EEPS
24 instrument is capable of generating reproducible total particle concentrations.

25 The overall average percent difference between the EEPS and UCPC particle number
26 concentrations was between 19% and 20%, with UCPC always higher. This was consistent
27 regardless of particle composition and for both stable and varying particle concentrations. The
28 instrument percent difference, which varied 5% to 25% during individual tests, results from the
29 different measuring techniques of the two instruments and appears to be a consequence of the
30 EEPS under-predicting concentration changes for the tested number concentrations

31 Constant and transient dilution factor tests were performed to compare the instruments.
32 The potentiometer on the ASET, which controls stage-2 dilution, was always kept at 7.1.
33 Constant concentration tests were performed step-wise by adjusting the MD19-2E potentiometer
34 in increments of 10, beginning at 10% and ending at 100%. The potentiometer was kept at each
35 increment for a minimum of 60 seconds. Transient tests were performed by rapidly changing the
36 MD19-2E potentiometer to random values and was kept between 10 and 100%. FIGURE 1 shows
37 typical results from both steady and transient concentrations.
38

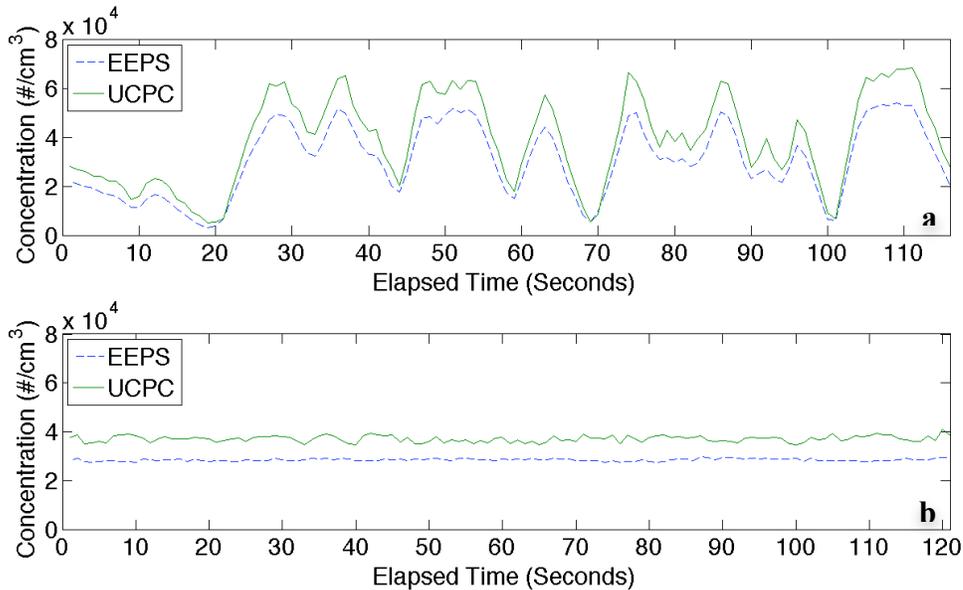


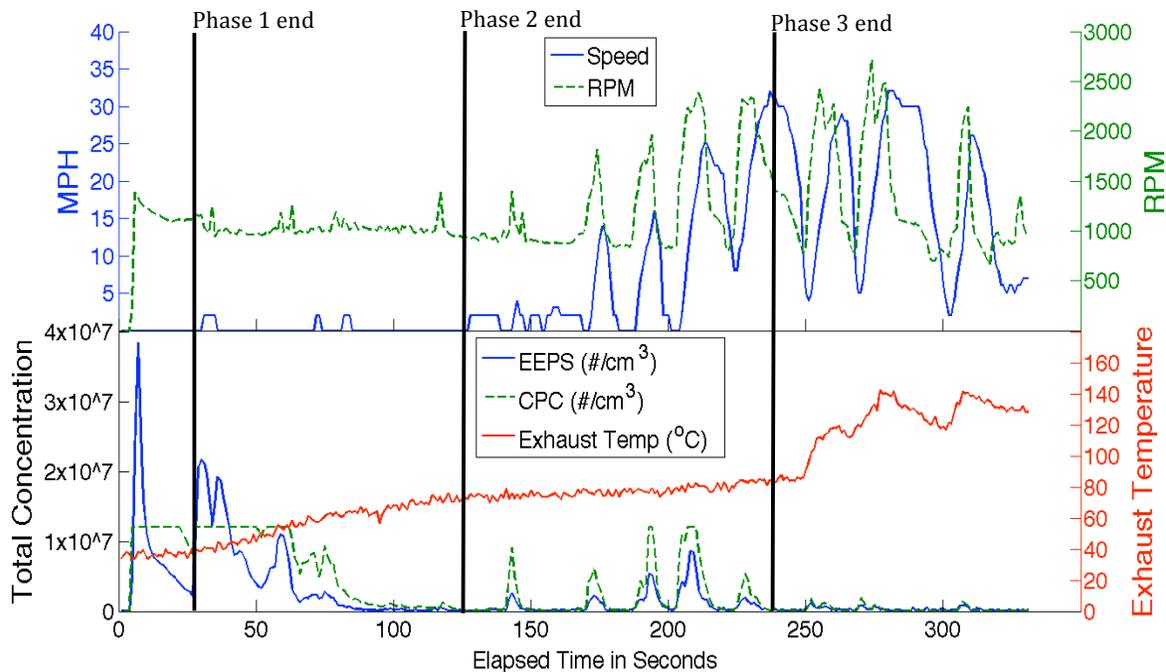
FIGURE 1 Tests to compare the EEPS and UCPC under a) transient concentrations and b) constant concentration of laboratory-generated sodium chloride particles.

The instrument response of both the EEPS and UCPC were very consistent. Comparing the EEPS to the UCPC, the two instruments track very well together with nearly identical response curves. From FIGURE 1, the EEPS seems to under-predict concentration changes when compared to the UCPC for these laboratory test conditions. However, neither instrument appears to show a hysteresis effect. Under low concentrations (<5,000 particles per cm³), the EEPS was much more variable, and therefore less accurate, due to electrometer noise. Emissions concentrations below 5000 particles per cm³ should therefore be interpreted with caution. The EEPS and UCPC had a consistent concentration difference, which is evident from the lines appearing to get closer under lower concentrations and farther apart under higher concentrations. The average percent difference between the two instruments remains near 19%.

The UCPC max concentration limit of 99,900 is a major drawback when looking at engine emissions. Even with a high dilution ratio (1:120), the UCPC reached its maximum limit during much of the cold start. Using a higher dilution ratio will cause lower concentrations to be undetected by the EEPS. For these reasons, TOTEMS employs both particle instruments.

Defining Cold Starts

Although previous studies used catalyst light-off [18] or engine coolant temperature [12] to characterize cold starts, this study used particle number emissions behavior because of patterns observed in the data. Total cold start duration was characterized by 3 separate Phases based on significant changes in particulate emissions behavior. FIGURE 2 shows these different Phases, and the end of each phase is labeled (note the flat tops to the UCPC curve in FIGURE 2 resulted when maximum concentration limits were reached).



1
 2 **FIGURE 2 Total concentrations from the EEPS and UCPC during run 1, corrected for**
 3 **dilution (bottom panel). RPM, speed and exhaust temperature are included for**
 4 **reference. Three cold start phases are shown by vertical lines. Concentrations are in**
 5 **#/cm³ and exhaust temperature is °C.**
 6

7 Cold start emissions were defined to begin when the particle number concentration increased to 3
 8 times greater than the tunnel blank background of 500 to 800 particles/cm³ [EEPS] and 70 to 100
 9 particles/cm³ [UCPC] (FIGURE 2 at 4-seconds). Note that reported particle number
 10 concentrations were not corrected for losses in the transfer lines. Phase 1 of the cold starts
 11 occurred when the car was in park and the engine running; it is characterized by the highest total
 12 particle number emissions of each run. Phase 1 began at engine-on and depended on when the
 13 driver put the car in gear. Although not ideal for analysis, this is consistent with the variability
 14 associated with real-world driving.

15 Phase 2 of the cold start was characterized by vehicle operation – starting when the
 16 automatic transmission was shifted to “Drive” (FIGURE 2 at 27-seconds). The end of Phase 2
 17 was defined by emissions behavior – when Phase 2 idle particle concentrations dropped to levels
 18 comparable to “hot-stabilized” idle concentrations. The vehicle was moving during Phase 2, but
 19 predominately at idle-creep (<5 miles per hour) in a parking lot setting.

20 Phase 3 was characterized by a particle emissions pattern showing elevated emissions
 21 during acceleration events (FIGURE 2 127-237 seconds). The end of Phase 3 occurred when
 22 particle concentrations satisfied three criteria: (1) total particle number concentrations were
 23 comparable to hot-stabilized idle (equivalent to ~2 times above EEPS instrument blank [or 1400
 24 #/cm³] and ~250 times above UCPC instrument blank [or 100 #/cm³], prior to accounting for
 25 dilution), where hot-stabilized idle concentrations were determined from traffic signal idling data
 26 collected later in the full driving route; (2) the smallest particles (5.6 to ~10 nm) in the measured
 27 EEPS distribution are no longer detectable; and (3) there were no longer significant increases in
 28 total number concentration as a result of RPM increases. Therefore, Phase 1 was driver behavior
 29 dependent, while Phases 2 and 3 were characterized by particle number emissions patterns.

1 **EXPERIMENTAL RESULTS**

2
3 **Cold Start Emissions Analysis**

4
5 The broader measuring range of the UCPC at both the large and small particle diameter
6 ends (3 – 3000 nm [UCPC] vs. 6 – 560 nm [EEPS]) seems to indicate there are large numbers of
7 particles formed during cold start that are outside the EEPS’s measurement range during cold
8 start. This observation is drawn from the significant differences in particle number
9 concentrations (at least a factor of 3) that are seen only during the cold start (FIGURE 2) but
10 were not present while driving after the catalyst is hot or in the lab under controlled particle
11 generation. The difference can most likely be attributed to the nucleation of small particles below
12 the EEPS lower diameter limit of 5.6 nm, during engine start. Significant particle numbers larger
13 than 562 nanometers (EEPS maximum size) are unlikely in vehicle exhaust.

14
15 *Cold Start Total Particle Number Emissions Data*

16
17 Cold start data were collected on five separate days with the ambient temperature range from
18 20°C to 37°C for all starts. TABLE 1 summarizes the cold start data.

19
20 **TABLE 1 Summary of Cold Start Emissions Data***

Run No.	Cold Start Duration (seconds)	EEPS Total Cold Start (#/cm ³)	CPC Total Cold Start (#/cm ³)	Ave T (°C)	Ave RH (%)
1	230	7.11E+08	1.09E+09 **	20	28
2	165	3.50E+08	6.68E+08 **	37	23
3	191	7.23E+08	Instrument Malfunction	25	42
4	201	5.13E+08	8.35E+08 **	33	39
5	199	4.97E+08	7.96E+08 **	32	37

21
22 * EEPS and UCPC data are the sum of particle number concentrations for the full cold start duration (Phases
23 1-3 inclusive).

24 ** Indicates the instrument reached its maximum concentration limit (9.99x10⁴) during the cold start period,
25 meaning data is biased low.

26
27 The actual total concentration difference between the EEPS and UCPC was even greater than
28 documented in TABLE 1 because the UCPC reached its upper measurement range (9.99x10⁴) for
29 extended periods during the cold start, whereas the EEPS never reached its maximum limits
30 (Refer to FIGURE 2). Cold start duration presented in TABLE 1 was determined from engine-on
31 to the end of Phase 3.

32
33 *Cold Start Phases: Duration and Particle Modes*

34
35 FIGURE 3 shows the contour plot of EEPS number distributions over the full duration of one
36 cold start with corresponding vehicle speed, RPM and exhaust temperature. Data is presented as
37 dN/dlogDp, which denotes the particle number concentration normalized to the individual
38 diameter range on a given EEPS channel (or size bin), (i.e. channel 1 is from 5.6 to 6.4 nm but
39 channel 32 is from 487 to 562 nm) Cold start Phase 3 clearly shows particle concentrations

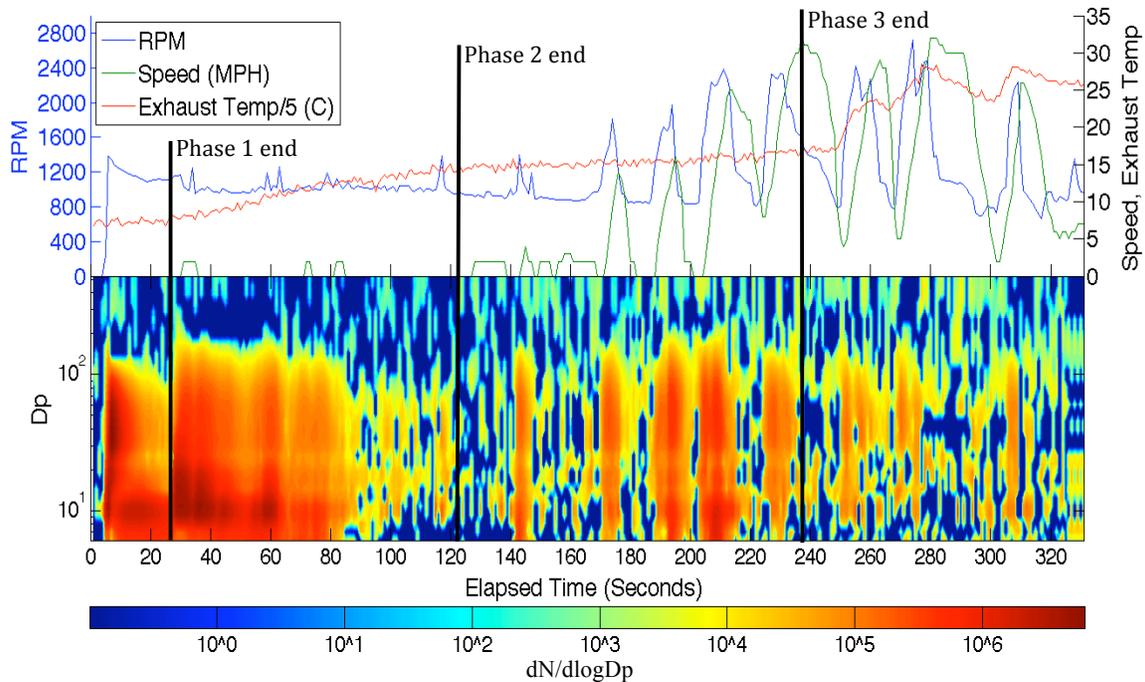
1 spiking with increased RPM – which, in this case, indicates acceleration – for all particle sizes
2 below 154 nm. This effect diminished as exhaust temperature gradually increased (see FIGURE
3 2, 140 to 237 seconds, for example).

4 Phase 1 lasted for 15 to 35 seconds and the highest measured concentrations were for
5 particles between 8 and 60 nanometers (nm). Particles larger than 20 nm diameter had extremely
6 high concentrations for only about 5 seconds after engine start before dropping significantly; this
7 behavior can most likely be attributed to enrichment at engine start [9]. Even though the larger
8 (>20 nm) particle concentrations dropped quickly in phase 1, their concentrations remained as
9 much as 100 times higher than hot-stabilized idle for the duration of Phase 1.

10 Phase 2 lasted 60 to 95 seconds with significant increases in number concentrations for
11 particles 5.6 to 200 nm in diameter. The highest concentrations, however, occurred for 10 to 30
12 nm particles and were 2 orders of magnitude above hot-stabilized idle concentrations. Increases
13 in RPM, which indicate vehicle acceleration, during Phase 2 resulted in significant particle
14 concentration – and subsequently emission rate – increases with particle diameter centered
15 around 10 nm.

16 Phase 3 had variable particle number emissions that depended on driver behavior in
17 urban traffic and lasted for 90 to 140 seconds. During Phase 3, when at idle (~700 RPM), the
18 concentrations remained comparable to hot-stabilized idle concentrations. When RPM increased,
19 which again indicates acceleration, 6 to ~154 nm particle concentrations increased with the
20 highest concentrations at 10 and 30 nm. This is largely because the catalyst is unable to
21 efficiently remove particle precursor vapors during Phase 3 due to the catalyst being below its
22 optimal operating temperature and decreased residence time. The particle number increase for
23 sizes between 6 and ~154 nm occurred even under engine loads of 30% to 40% and acceleration
24 of 2 to 3 mph/sec. The emission of Phase 3 particles can likely be attributed to enrichment during
25 acceleration (>30 nm) events [9], mechanical release of particles (>30 nm) and increased
26 flowrates (<30 nm). As exhaust flowrate increases, the catalytic converter becomes more
27 inefficient due to reaction kinetics and shorter residence time in the catalyst. Although
28 acceleration rates during the warm-up driving were relatively low (2 to 3 mph/sec), even small
29 increases in flowrate may be enough to cause significant increases in particulate number
30 emissions if the catalytic converter is not operating at its optimal level. This may therefore
31 indicate a gradual increase in the efficiency of the catalyst relative to its temperature until the
32 optimal temperature is reached. The larger particles (>60 nm) observed throughout the cold start
33 may, in part, be attributed to re-entrainment from the mechanical release of solid particles from
34 the exhaust system caused by increased flowrates. At the end of Phase 3 (which is also the end of
35 the cold start at 237 seconds on FIGURE 3), the smallest particles measured by the EEPS (5.6 to
36 ~10 nm) are no longer detectable or have very low concentrations (see FIGURE 3, 237 to 330
37 seconds), indicating the catalyst is efficiently removing compounds that potentially nucleate to
38 particles.

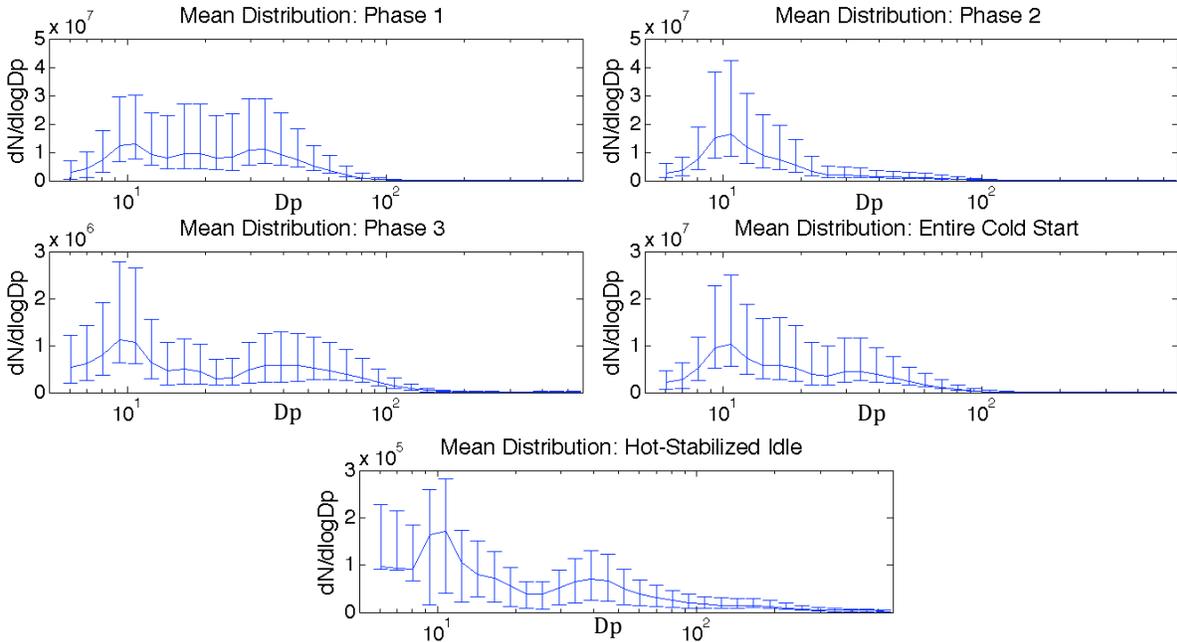
39 As expected, the highest particle number concentrations occurred for particles in the
40 nanoparticle ($D_p < 50$ nm) range, which accounted for over 97% of the particles over the entire
41 cold start (Phases 1-3).



1
2
3 **FIGURE 3** Contour plot of particle number distribution (bottom) over first 330 sec of
4 **Run 1** test period. Upper plot is time-aligned data on engine speed, vehicle speed and
5 **exhaust temperature**. Exhaust temperature was divided by 5 to achieve same scale as
6 **vehicle speed**. Engine on at 4 seconds, vehicle in gear at 27 seconds and moved only at idle-
7 **creep speed** out of parking area. Vehicle accelerated onto Colchester Avenue at 141
8 **seconds**.

9
10 *Mean Particle Size Distribution*

11
12 EEPS data for each phase of the cold start over the 5 runs were averaged to evaluate changes in
13 particle distribution over time after engine ignition. The particle number distributions changed
14 consistently between phases of the cold start (FIGURE 4), but the particle diameter of the
15 distribution “modes” was consistent run-to-run. The “modes” are the particle diameters of
16 the distribution’s peaks. Modes are typically assigned as sequential numbers based on
17 particle diameter (x-axis), not the concentration (y-axis). In Figure 4, mode 1 (10 nm) had
18 the highest concentration, mode 2 (18 nm) had the lowest concentration, and mode 3 (50
19 nm) was intermediate in concentration.
20



1
 2 **FIGURE 4 Cold Start Mean Particle Number Distributions. Note the different y-scales.**
 3 **The solid line represents mean EEPS concentrations over 5 cold start runs and the error**
 4 **bars represent the overall range from 5 runs. The distribution under hot-stabilized idle**
 5 **operation is included for reference (note it has much lower concentrations and a chiefly**
 6 **bimodal distribution).**

7
 8 Mode 1 (10 nm) was present in all three phases of the cold start, but mode 2 (18nm) and mode 3
 9 (50 nm) were distinct only for Phases 1 and 3. Phase 1 had the highest variability in particle
 10 number for all particle diameters. During hot-stabilized idle operation (FIGURE 4, bottom), the
 11 greatest variability and highest emissions were for sizes between 8.65 and 11.55 nm,
 12 corresponding to Mode 1 of the cold start distributions. Like Phase 2, hot-stabilized idle
 13 operation lacked a distinct mode 2 (18nm)..

14
 15 *Magnitude and Duration of Different Particle Sizes During Cold Start*

16
 17 Concentrations from the EEPS were aggregated to 10 bins based on similar trends during cold
 18 start. EEPS sizing ranges are equally spaced on a log-scale, so the smallest ranges – and
 19 therefore the best resolution – are in the smallest particle sizes. TABLE 2 shows the different
 20 aggregate bin size ranges, the average particle concentration (magnitude) from all combined runs
 21 and the temporal duration over which that particle size was detectable for each phase of the cold
 22 start. Concentrations from hot-stabilized idle operation are also provided for comparison. The
 23 duration of particles for each size range, except for particles from 5.6 to 7.5 nm, began at the
 24 beginning of each Phase and lasted, on average, the length of time provided in TABLE 2. Particle
 25 sizes from 5.6 to 7.5 nm demonstrated slightly different behavior from all other sizes and
 26 therefore comprised the first aggregated bin. Particle number did not increase significantly after
 27 engine start for 3 to 5 seconds, then increased to the levels reported in TABLE 2. The particle
 28 number concentrations in bin 1 also changed very little through Phases 1 and 2, whereas sizes
 29 greater than 7.5 nm showed a clear end to Phase 1. Particle sizes between 7.5 and 100 nm
 30 showed very similar patterns in time series concentration, so these sizes were aggregated to 8

bins based on the duration the particles were detectable. For example, in TABLE 2, bin 7.5 to 10 nm and bin 10 to 13.3 nm had nearly identical mean duration over each cold start Phase. During the aggregation process, particle sizes from the EEPS 32 size ranges demonstrating these similarities were combined. Larger particle size concentrations (i.e., diameters 100 to 562 nm), even when aggregated, were always at least 2 times (only representing ~1% of total number emissions) lower than other aggregated particle concentrations, so a more disaggregate approach was deemed unnecessary for this 10th bin.

TABLE 2 Mean particle number magnitude (#/cm³) and duration (seconds) for the three cold start phases and hot-stabilized idle for aggregated particle size ranges.

Particle Size Range	Phase 1			Phase 2			Phase 3			Hot-Stabilized Idle
	Magnitude	Duration	StDev of Duration	Magnitude	Duration	StDev of Duration	Magnitude	Duration	StDev of Duration	Magnitude
5.6 - 7.5	4.52E+05	24	6.56	3.83E+05	43	28.02	6.78E+04	91	36.06	7.20E+03
7.5 - 10	1.23E+06	27	4.73	1.43E+06	42	27.54	1.14E+05	91	37.15	1.31E+04
10 - 13.3	1.38E+06	27	4.73	1.76E+06	42	27.54	1.01E+05	93	27.41	1.72E+04
13.3 - 17.8	1.10E+06	20	5.13	1.01E+06	42	27.54	5.56E+04	97	31.33	1.30E+04
17.8 - 23.7	1.09E+06	20	5.13	5.48E+05	27	26.10	4.26E+04	97	31.33	6.93E+03
23.7 - 31.6	1.19E+06	15	2.65	2.50E+05	25	28.02	4.66E+04	92	24.01	9.31E+03
31.6 - 48.7	1.74E+06	16	2.73	2.85E+05	25	28.02	1.05E+05	97	15.28	2.69E+04
48.7 - 75	6.62E+05	12	2.31	1.87E+05	27	26.10	8.30E+04	100	36.06	1.84E+04
75 - 100	7.79E+04	6	1.73	5.91E+04	20	10.41	3.04E+04	60	52.92	6.38E+03
100 - 562.3	2.05E+04	2	1.15	3.49E+04	7	2.52	1.93E+04	40	57.74	7.75E+03

Phase 3 was different than the first two Phases due to the differing behavior during on-road driving (variable RPM, vehicle speed and exhaust flowrate). Particle number increased for most particle sizes up to 154 nm (as shown in FIGURE 3) when RPM and flowrate increased, so the average durations of most aggregate size bins were quite similar in Phase 3. Also, Phase 3 concentrations were 1.6 to 17 times lower (depending on particle size) than Phases 1 and 2. This was a direct result of having generally low emissions during many episodes of idle operation from on-road travel under urban driving conditions. As was also indicated by [12], mean particle size decreased through each Phase, which is shown in TABLE 2 by smaller particles having longer durations. Lastly, the variability in cold start duration, and subsequently total cold start number emissions, is evident in the standard deviation. The standard deviation was typically larger for the smaller particles, indicating greater variability from day to day in the time period over which the smallest particles were emitted.

Effect of Ambient Temperature

Although all data collected to date were for relatively warm days, with ambient temperatures between 20 and 37°C (refer to TABLE 1 for ambient conditions), a trend was observed between ambient temperature and total cold start duration. The maximum particle number concentrations emitted, as measured by the EEPS (the UCPC always reached its maximum concentration limit), were not influenced by the ambient temperature over the range of ambient temperatures sampled, but the sum of particles emitted throughout the entire duration of the cold start clearly indicated that, because of increased cold start duration, cooler days resulted in greater total cold start particle number emissions. The EEPS total particle number was 2 times higher for a 20°C day versus a 37°C day. However, despite a shorter duration, the 25°C day cold start emissions were slightly higher than the 20°C day by 2%. This is the only inconsistent day from the general trend.

1 The UCPC was tentatively quantified, but because it reached its maximum limit during much of
2 the cold start, its concentrations are biased low. The UCPC recorded the 20°C day as 1.6 times
3 higher than the 37°C day. The observed effect of ambient temperature on cold start emissions is
4 consistent with previous work [9, 12, 19], including studies that only quantified gaseous
5 emissions such as HC and CO [17, 18], which are precursors to particulate emissions [9].
6
7

8 **CONCLUSIONS**

9

10 Although this dataset was small, based on only one vehicle, and measurements were performed
11 with ambient temperature above 20°C, the value of second-by-second particle number
12 distributions, which is rarely available in previous studies, is documented. High temporal
13 resolution of the EEPS instrument allowed identification and quantification of shifting
14 distributions between phases of the cold start. This means a slower Scanning Mobility Particle
15 Sizer instrument is inadequate for fully understanding particle behavior during the entire cold
16 start. Also, all cold starts were performed with ambient temperatures above 20°C. Although there
17 was clearly a relationship between cold start duration and magnitude and ambient temperature,
18 more data needs to be collected in fall and winter months to determine how great of an influence
19 ambient temperature has on particle number distributions.

20 It's extremely important to understand the distribution, as well as the magnitude and
21 duration of cold starts. Most exhaust particles (~99%) are less than 100 nm in diameter. For the
22 minivan tests, 3 modes were present at 10 nm, 18 nm and 50 nm with the highest concentrations
23 and longest duration occurring at 10 nanometers. Cold start duration was greatly influenced by
24 ambient temperature ranging from 165 to 230 seconds, where concentrations measured by the
25 EEPS were at least 10 times and most 100 times greater than hot-idle concentrations. If the
26 UCPC is considered, cold start particle number concentrations were at times at least 3 orders of
27 magnitude greater than hot-stabilized idle concentrations. This has implications for better
28 understanding the health affects associated with particle number emissions from vehicles.

29 Instrument capabilities currently have serious limitations when quantifying vehicle
30 particle number emissions. From this study, it is seen that the full particle number distribution
31 cannot be collected from light-duty vehicles with the necessary particle size resolution using any
32 single commercially available instrument. The increased difference between the EEPS and
33 UCPC occurring during the cold start indicated significant numbers of particles outside the EEPS
34 measuring range. Therefore, comprehensive, multi-instrument studies must be conducted
35 representative of vehicle types, fuels, maintenance levels, emissions control devices, vehicle age,
36 and varying weather conditions to better understand cold start emissions.
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